

MIGUEL ÂNGELO PARDAL  
JOÃO CARLOS MARQUES  
MANUEL AUGUSTO GRAÇA  
Scientific Editors

# Aquatic Ecology of the Mondego River Basin Global Importance of Local Experience



Coimbra • Imprensa da Universidade

MIGUEL ÂNGELO PARDAL  
JOÃO CARLOS MARQUES  
MANUEL AUGUSTO GRAÇA  
Scientific Editors

# Aquatic Ecology of the Mondego River Basin Global Importance of Local Experience



Coimbra • Imprensa da Universidade

COORDENAÇÃO EDITORIAL  
Imprensa da Universidade de Coimbra

CONCEPÇÃO GRÁFICA  
António Barros

INFOGRAFIA  
António Resende  
Estímulus [design] • Coimbra

EXECUÇÃO GRÁFICA  
GRAFIASA

ILUSTRAÇÃO DA CAPA  
P. P. Cunha e J. Dinis

ISBN  
972-8704-04-6

DEPÓSITO LEGAL  
175038/02

© JANEIRO 2002, IMPRENSA DA UNIVERSIDADE DE COIMBRA



OBRA PUBLICADA COM O PATROCÍNIO DE:  
IMAR – INSTITUTO DO MAR  
IPIMAR – INSTITUTO DE INVESTIGAÇÃO DAS PESCAS E DO MAR



TIAGO MÚRIAS <sup>1</sup>  
JOÃO CARLOS MARQUES <sup>1</sup>  
JOHN GOSS-CUSTARD <sup>2</sup>

## EFFECTS OF HABITAT LOSS ON WADERS (AVES, CHARADRII) IN THE MONDEGO ESTUARY: A SYNTHESIS AND FUTURE PERSPECTIVES

### Abstract

The destruction of the salines represents one of the major threats to the wader populations in the Mondego estuary (Portugal). Although the actual levels of competition in the intertidal habitats are low, thus allowing for many displaced birds from the salines to settle there, these birds will still lose part of their previous feeding space and time. The ongoing eutrophication process in the estuary will further aggravate this. The conservation of this area, in what the wader populations are concerned, demands an integrated policy of all authorities and individuals involved.

### Introduction

Habitat loss is probably the most intensively examined of the factors that are known to potentially affect the survival of estuarine wader populations (Goss-Custard et al. 1996a, b). Most studied cases have focused on the consequences for waders of the direct loss of intertidal low-water feeding areas through land reclamation for industrial, agricultural or water storage purposes (Davidson et al. 1991), and of indirect losses due to the submersion of previously accessible feeding areas following the construction of tidal power and storm-surge barriers or sea-walls, as a consequence of a sea-level raise (Meire et al. 1994, Lambeck et al. 1996). In contrast, there have been few studies dealing with the consequences of the loss of supratidal feeding areas, probably because in north European estuaries these are not heavily used, or critically endangered, although their importance for waders is recognised (Davidson and Evans 1986, Hötker 1994)

The situation may be different in the south European Atlantic estuaries. Supratidal habitats, such as the salines, usually occupy large areas within the estuaries and "nas".

---

<sup>(1)</sup> IMAR – Instituto do Mar, Centro Interdisciplinar de Coimbra a/c Departamento de Zoologia, Universidade de Coimbra, 3004-517 Coimbra, Portugal

<sup>(2)</sup> Centre for Ecology and Hydrology (CEH) - Dorset, Dorset DT2 8ZD, United Kingdom

and seem to be intensively used throughout the tidal cycle, by a lot of species (Rufino et al. 1984, Perez-Hurtado et al. 1991, 1993b). These man-made wetlands are currently more threatened in the southern European estuaries than are the natural intertidal areas, mainly due to their abandonment and/or transformation into fish-farms or rice fields (Rufino and Neves 1992, Neves and Rufino 1995, Perez-Hurtado and Hortas 1993a).

The loss of salines would affect the two broad groups of birds that use them in slightly different ways. One group of waders feed over the low-water period on the intertidal mudflats. If these birds are, however, unable to obtain all they require, they then feed in the salines over high tide when their preferred intertidal flats are no longer available. These birds were called the "intertidal birds". The other group of birds is those which feed in the salines throughout the tidal cycle, both at low and high water, and rarely if ever utilise the mudflats of the intertidal area. These were called the "salines birds" (Múrias 1997).

The removal of the salines would remove all the feeding space presently used by the "salines birds", which would thus lose all the current feeding time, as well. Those birds would probably try to settle in the mudflats (Meire et al. 1994, Lambeck et al. 1996). Their survival in the estuary would then depend on their actual feeding requirements in the salines (i.e. if they are able to get all the food they need within the 8.5 hour limit for feeding at low-water in the mudflats), and of the present level of competition in the intertidal areas (Goss-Custard and West 1997).

The "intertidal birds", on the other hand, would lose the extra feeding time they presently use at high-water, when their main feeding areas in the intertidal mudflats are unavailable. On many occasions, particularly in the more energetically-demanding periods of the year, these birds will not be able to recover this feeding time by extending their foraging at low-water through an increase in the foraging time and/or in the intake rate (Davidson and Evans 1986, Goss-Custard et al. 1996c). Their only chance, if they were to remain on the area would be, therefore, to look for alternative supratidal sites.

The present paper summarises the results of a research study conducted on the problem and potential consequences for waders of habitat loss in the estuary of Mondego, in 1993-95 (Múrias 1997). Three main issues were examined in this study of the potential effects of habitat loss on waders in the Mondego estuary: a) the importance of the salines as feeding habitats; b) the ability of the mudflats to support the displaced birds in case the salines were destroyed, and c) the effects on all this, of the increasing eutrophication process which is actually taking place in the Mondego estuary.

## Study Area and Methods

The Mondego is a warm-temperate estuary in a region with a basic Mediterranean climate. The terminal part of the estuary consists of two arms, north and south, that surround an alluvion-formed island, the Morraceira (Figure 1). Due to

distinct hydrodynamics in the two arms, the south arm is heavily silted-up, thus providing the most important intertidal areas (134 ha) which, along with the salines of the Morraceira Island, are the most suitable feeding areas for waders in the estuary. There is a north-south gradient of increasingly finer sediments in this arm (Marques et al. 1993). In late spring and summer of some years, the finer sediments are covered with extensive weed mats, which remain in place until the beginning of the winter (Múrias 1997). For the purpose of this study, the intertidal area was subdivided in three naturally delimited sub-areas, of 66.8 ha (upstream section), 30.9 ha (middle section) and 36.1 ha (downstream section).

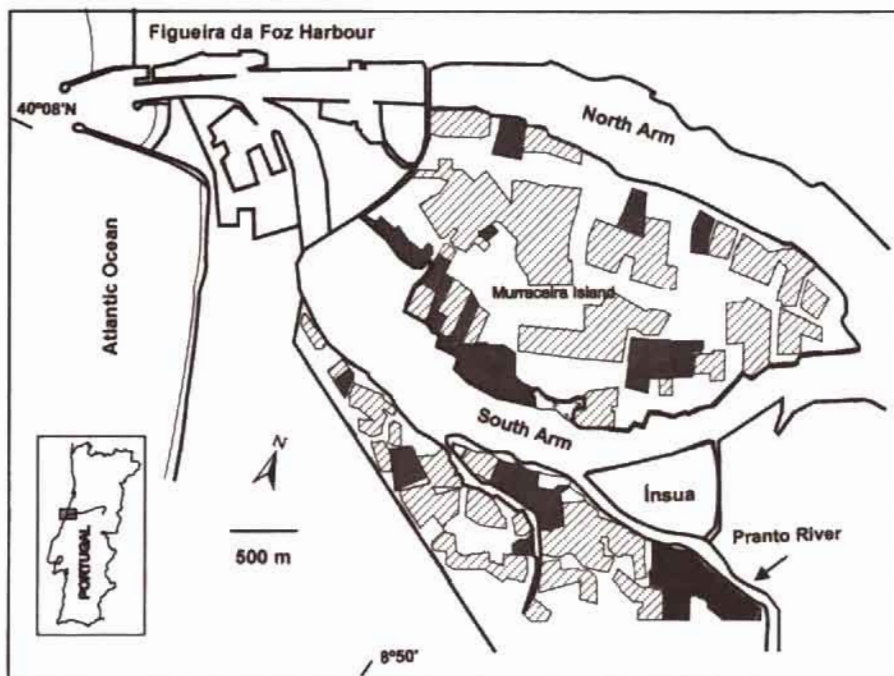


Fig. 1. The estuary of Mondego showing the location of the main area of artisanal salines (stippled), fish-farms (dark-grey) and industrial salines (black). The abandoned salines of the north arm and the Insua are not represented. The main intertidal areas are located along the south arm

The basic method used to estimate bird numbers were high-water and low-water censuses, carried out with 10 × 50 binoculars and a 20-60 × 50 telescope, in the Morraceira's salines and in three fixed plots along the south arm of the estuary. The counts were performed each month, from October 1993 to January 1994 and fortnightly onwards, to May 1995.

Invertebrate samples were taken along fixed transects in the south arm with a corer, 95 cm<sup>2</sup>, 5 cm deep, taken to the laboratory and sorted, identified, counted and measured according to the usual procedures (Marques et al. 1993a, Múrias 1997, Múrias et al. 2002).

A more detailed account of the methods used can be found in Múrias et al. (1996) and Múrias (1997).

## Results and Discussion

### Loss of salines

Overall, the results of this study suggest that, due to time loss, many birds in this estuary could be in trouble were all the salines to be destroyed.

The salines were shown to be intensively used by most wader species at high-water, both as complementary feeding areas to the main feeding places in the intertidal mudflats, but also at low-water, as alternative feeding areas, with a high proportion feeding there at all times (Table 1). However, the number of salines available to the waders in the estuary is rapidly decreasing, due to abandonment and transformation of the ponds into fish farms (Table 2).

Table 1. Percentage of the total number of birds of each species counted in the whole study area at low-water (N) that used the Morradeira's salinas at low and high-water (in parenthesis, percentage of birds that were feeding). Values represent the average number of birds present per count in the estuary over the whole study period (October 1994 to May 1995, maximum n=38 counts).

	N ± SE	(n)	Low-water %	High-water %
Black-winged Stilt <i>Himantopus himantopus</i>	62.4 ± 9.9	(22)	79.7 (68.9)	87.4 (70.0)
Redshank <i>Tringa totanus</i>	12.4 ± 3.7	(26)	53.3 (65.0)	69.0 (67.4)
Little Stint <i>Calidris minuta</i>	25.3 ± 11.1	(21)	52.9 (59.3)	77.4 (59.3)
Dunlin <i>Calidris alpina</i>	451.6 ± 74.4	(37)	20.0 (70.1)	60.0 (70.2)
Kentish Plover <i>Charadrius alexandrinus</i>	91.5 ± 10.6	(38)	18.8 (48.3)	59.4 (42.4)
Ringed Plover <i>Charadrius hiaticula</i>	66.7 ± 11.3	(34)	12.1 (61.5)	57.7 (50.7)
Whimbrel <i>Numenius phaeopus</i>	7.9 ± 3.4	(7)	6.9 (5.0)	25.5 (3.6)
Grey Plover <i>Pluvialis squatarola</i>	86.6 ± 10.8	(28)	1.3 (3.8)	14.3 (8.4)
Avocet <i>Recurvirostra avosetta</i>	388.6 ± 53.4	(18)	< 0.1 (0.5)	0.1 (3.7)

**Note:** Only the species with 5 birds per season and per year, or more, were considered



Table 2. The loss of salines by abandonment and transformation into fish-farms in the Morraceira Island up until 1994. Total number of salines in the Morraceira before 1984 was 229, with an area of 305.1 ha. These figures do not include some salines levelled prior to 1984.

	1955 - 1984	1984 - 1994	Total
Number of salines abandoned	26	30	56
Number of salinas transformed	-	22	22
<b>Total number lost</b>	<b>26</b>	<b>52</b>	<b>78</b>
% of total number lost	11.4 <sup>1</sup>	25.6 <sup>1</sup>	34.0
Rate of loss (salines.year <sup>-1</sup> ) <sup>2</sup>	0.9	5.2	1.9
Area abandoned (ha)	35.3	51.0	86.3
Area transformed (ha)	-	19.7	19.7
<b>Total area lost (ha)</b>	<b>35.3</b>	<b>70.7</b>	<b>106.0</b>
% of total area lost	11.6 <sup>2</sup>	26.2 <sup>2</sup>	34.7
Rate of loss (area.year <sup>-1</sup> ) <sup>3</sup>	1.2	7.1	2.5

<sup>1</sup> Calculated in each period by excluding the total number already lost in the preceding period(s).

<sup>2</sup> Calculated in each period by excluding the total area already lost in the preceding period(s).

<sup>3</sup> The number of years is: pre-1984 - 30 years; 1984-94 - 10 years.

Ultimately, this process may force these birds to leave the salines and try to re-establish themselves on the mudflats, if they are still able to accommodate them.

All depends on the present-day level of competition in the mudflats and how close the densities of the birds are, in these areas, to reach the point where their mortality rates, or body condition, become density-dependent. As stressed by Goss-Custard and West (1997), the moment were this point is reached is what really matters to this issue, as it ultimately leads to a reduction in local bird numbers, through intensified competition, well before the carrying capacity is reached, if ever.

At the moment, all data seems to suggest that the effects of competition on the mudflats, either through interference or through resource depletion, are weak (Figure 2). Therefore, a number of the birds that use the salines as alternative feeding areas to the mudflats could eventually re-establish themselves in the intertidal flats at low-water, buffering the loss of low-tide feeding space in the salines. However, even these birds seemed to intensively use the salines for feeding at high-water (Table 1). So, the need for finding supplementary supratidal feeding areas could involve both groups of birds, the mudflat-feeders and the salines-feeders.

The problem arises because no other supratidal habitat in the estuary is likely to offer the good feeding conditions that birds can presently find in the salines. The rice fields are only usable at certain times of the year, at other times being too flooded (in winter) or too dry (in summer). Furthermore, it is unlikely that birds that do not feed on mudflats covered by algae the "bare-sediment" specialists, such as the small plovers (Kentish and Ringed Plovers) would use this habitat at all. Elsewhere, saltmarshes are also commonly used supplementary feeding sites for mudflat feeders, both in Europe

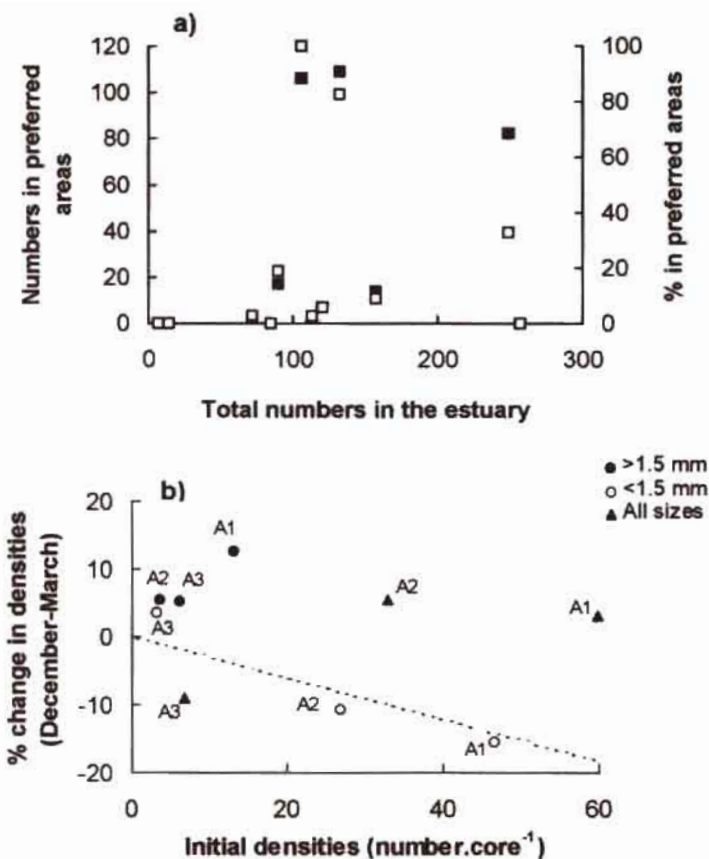


Fig. 2. Graphics showing the absence of competition by (a) interference of feeding birds and (b) resource depletion. In (a) it is illustrated the evolution in bird numbers (■) and in the percentage of these birds in preferred areas (□) for Ringed Plover *Charadrius hiaticula*. The preferred areas are defined as being those areas firstly occupied by birds (Goss-Custard, 1980). In case interference competition was operating, the number of birds in preferred areas should initially increase and then, with the increase in total numbers in the estuary, reach a plateau, indicating that these areas were already full. The percentage of birds in preferred areas, on the contrary, should follow the opposite trends, as more and more birds were driven out from the preferred areas by competition and forced to establish in poorer areas. In (b) is investigated whether the depletion of *Hydrobia ulvae* populations by birds in the estuary during the winter (November-March) was density-dependent (indicated by the dotted line), thus suggesting that the competition for food was high (birds tend to concentrate where the densities of their prey are higher). No such trend was found for the class of prey consumed by the birds (> 1.5 mm), therefore suggesting that no significant depletion occurred during the study period.

and in South Africa (Davidson and Evans 1986, Velasquez et al. 1990, Velasquez and Hockey 1991). However, in the Mondego, they are mainly formed by *Spartina* sp., whose dense stands are known to deter many small species, such as Dunlin (Goss-Custard and Moser 1988). Only the larger species (Grey Plover, Whimbrel, godwits) were occasionally seen using this habitat in the Mondego, but these would probably be the least affected species by the loss of the salines (see Tables 1 and 3).

Table 3. The calculated amount of feeding space (expressed as the percentage of "bird feeding-hours", or BFH - number of feeding birds  $\times$  number of hours spent feeding) and feeding time lost by several wader species in the estuary of Mondego in case the salines are destroyed, and the predicted increase in the feeding pressure on the mudflats if all birds displaced from the salines try to settle there. Two groups of birds with slightly different strategies were considered: those that used the salines as alternative feeding to the mudflats, throughout the tidal cycle (the "salines birds"), and those that used the salines only at high-water, as supplementary feeding areas to the intertidal mudflats (the "mudflat birds"). Values are means for 1993-94 and 1994-95, averaged for autumn, winter and spring; therefore, N=6 for all species, except for Whimbrel, for which N=2, and Avocet, for which N=5.

	Total feeding (BFH $\pm$ 1SE)	"Mudflat"	birds	"Salinas" birds	Predicted increase in BFH in the mudflats (in %)
		Space loss (%)	Space loss <sup>1</sup> (%)	Space loss <sup>1</sup> (%)	
Redshank	199.9 $\pm$ 77.2	70.0	25.0	36.9	66.0
Little Stint	334.0 $\pm$ 154.3	68.0	-	18.5	90.0
Dunlin	2318.1 $\pm$ 1780.7	23.7	37.5	16.7	42.6
Kentish Plover	668.6 $\pm$ 71.4	22.1	9.4	16.5	34.8
Ringed Plover	610.6 $\pm$ 21.1	9.8	16.3	20.7	26.9
Grey Plover	449.4 $\pm$ 62.2	3.5	13.2	-*	4.5
Avocet	1256.0 $\pm$ 392.2	0.5	-	0.1	1.1
Whimbrel	56.9 $\pm$ 6.7	0.0	-	-*	0.5
<b>All species</b>	<b>736.9 <math>\pm</math> 272.3</b>	<b>24.7</b>	<b>20.3</b>	<b>15.5</b>	<b>33.3</b>

<sup>1</sup> BFH in the mudflats at low-water + salines at both high and low-water

<sup>2</sup> Calculated as the percentage of feeding time above the 8 hours available for feeding in the mudflats that is currently used by the mudflat birds in the salines at high-water. Maximum feeding time was assumed to be 12.5 hours each tidal cycle, and total feeding time per species were: Redshank - 10 hours; Little Stint - 4.4 hours; Dunlin - 11.0 hours; K. Plover - 8.8 hours; R. Plover - 9.3 hours; G. Plover - 9.1 hours; Avocet - 5.5 hours; Whimbrel - 7.8 hours

<sup>3</sup> Equivalent to the same amount of time loss, as these birds are assumed to feed in the salinas throughout the tidal cycle

\* Less than 0.1%

A less suitable but still usable alternative supratidal site to the salines could be provided by the fish-farms. They have the advantage of being encircled by wire fences, thus providing quiet and relatively safe places from attack by raptors. In fact, waders use them now as roosting sites. However, the ponds are usually too deep for waders to feed there even for the long-legged species, and in practice they are used only when they are periodically emptied for cleaning (Perez-Hurtado and Hortas, 1993a, b). In the Mondego, even this periodic usage seems to be infrequent (at least it was never observed during the present study), probably due to the depth and to the narrow dimensions of the ponds.

### Blooms of algae

Loss of intertidal habitat could also occur in the estuary, through the increasingly extensive "blooms" of green macroalgae that occur seasonally due to eutrophication (Marques et al. 1993a, b). The presence of contiguous and extensive algae mats would be expected to decrease the abundance of many prey species (see, e.g. Everett 1994),

and the subsequent recolonisation after the algal crash, from the small islets of unweeded areas that would remain, would be insufficient to replace the losses (Raffaelli et al. 1989, 1991). Waders could then be unable to find enough food to meet their daily requirements in the previously weeded areas.

In fact, although the effects of the eutrophication at the low levels of the trophic chain can be assessed in a matter of 2-3 years (e.g. Soulsby et al. 1982, Desprez et al. 1992, Everett 1994), its consequences for waders may take more time to be established. Subtle changes in the diet of some waders, when the populations of their main prey are affected, usually appear even before any change in numbers begin to be noted (Desprez et al. 1992). Only in a later stage of the process do waders respond by changing their feeding areas. In the Mondego, no indication was obtained that the waders changed feeding areas due to the presence of algae (see Múrias et al. *Chapter 5.2 of this book*). On the contrary, although circumstantial, there was evidence that some species (e.g. Dunlin) could even be attracted to the mats during, or soon after, the algal "bloom", had taken place in early spring.

Apart for some methodological reasons and the time-scale used (Múrias et al. *Chapter 5.2 of this book*), another factor may account for the lack of a relationship between bird numbers and algae in the Mondego. The prey most consumed by waders in this estuary seem to be mobile sediment-water interface feeders, such as some errant polychaetes and *Hydrobia ulvae* (Lopes et al. 1998), which may be favoured, at least in the earlier stages of the season, by the growth of algae (Soulsby et al. 1982, Everett 1994), thus providing enhanced, even though seasonally-limited, food resources for the waders.

#### Implications for the management of the system

In summary, this study suggested that the main threat to estuarine waders in the Mondego at present is the continuing destruction of the supratidal habitats. It seems that there is some buffering capacity on the mudflats to receive a number of displaced birds from the salinas. However, the lack of sufficient area in the supratidal habitats themselves that would enable birds to recover the feeding time that would be lost with the loss of the salines, could be detrimental to the populations of many species. In the long-term, the effects of the loss of supratidal habitat could be further aggravated if the eutrophication continues to increase at the present rate, thereby perhaps also reducing the available feeding space for waders at low-water, if the growth was severe enough.

From a conservation point of view, there is always the possibility of creating artificial supratidal habitats to replace those that are lost (Davidson and Evans 1986, 1987, Hötter 1994). There are, however, some limitations on the creation of adequate artificial supratidal wetlands. These are (i) the large areas that are required to allow for the settlement of all the displaced birds; (ii) the need to provide similar habitats to those destroyed, particularly in terms of their sediment types and invertebrate faunas, in order to attract the same species that were displaced; and (iii) the need to begin the work some years (2-3) in advance of the destruction of the primary habitat, due to

the time required to find an appropriate place, prepare the area and allow the settlement and growth of the invertebrate prey (Davidson and Evans 1987). Even so, there is no absolute assurance that waders will accept the new sites. Hötter (1994) showed that at least two of three artificially-created supratidal habitats in the German and Danish part of the Wadden Sea, that were constructed to compensate for the loss of reclaimed intertidal habitats, did contribute to the increase the number of bird species and densities in the area, but they did not fully compensate the losses due to land claims.

It seems that prevention is still the best way to avoid the more deleterious effects of habitat loss. In the Mondego, there may still be time to reverse the present trend. Many salines have been abandoned, but not yet transformed. Deserted salines are not completely unattractive to waders (Múrias et al. submitted), although they are far less used than the active ones. An effective and relatively inexpensive way of recovering these salines for waders would be to pay their owners to keep them clean and to maintain a permanently controlled water level. This would avoid the salines conversion or drainage. An alternative or, even better, complementary solution would be to impose strict rules on the construction of new fish farms in the estuary, by improving the design of the pond walls, in order to create areas of shallow water, as it was suggested by Rehfish (1994) for man-made brackish lagoons in England. This could allow even the smaller waders to use the ponds, although some care should be taken to prevent the access of piscivorous birds (e.g. herons).

Regarding eutrophication, any local intervention (e.g. by imposing some form of treatment of the urban, agriculture and fish-farm discharges to the estuary) would not be enough. It would be also necessary to control the urban and agricultural discharges along the whole lower river valley, in which the majority of rice fields and other extensively irrigated lands are located. This is a very difficult task, however, as it requires the involvement of many different official and private organisations.

Besides the obvious need to preserve the estuarine biodiversity and the health of the whole ecosystem, of which waders are an important component, there is another important reason why the quality of the habitat for waders should be maintained or even enhanced in this estuary. Small estuaries like the Mondego, with relatively low number of waders, as compared to the major estuaries of the East Atlantic Flyway (Smit and Piersma 1989), may act as "emergency" sites for some migrating or wintering birds. Emergency sites are areas where, in normal conditions, few birds land, but where, under adverse weather, they may stage in great numbers (Piersma, 1987 in Smit and Piersma 1989). This may prevent many birds from starving, avoiding the high mortality rates which otherwise would probably occur. Moreover, the real number of birds of all species that use the estuary of Mondego may have been underestimated, particularly during the migratory periods. As Smit and Piersma (1989) showed for a small Moroccan estuary (Sidi Moussa), the spring migration peak of 7000 birds underestimated by 3 times the real number of individual birds that crossed the area during a two-month period, as investigated by an intensive counting (3-5 days counts) and colour-marking program. If this is also the case in the Mondego, its perceived importance for waders would naturally increase still further.

## Acknowledgements

This work was supported by the Portuguese Research Board (JNICT), through grants FMRH/BD/331/93 and PRAXIS XXI/BD/5570/95 (TM). The authors are also indebted to all the colleagues that helped in the fieldwork, and to an anonymous referee for the constructive comments.

## References

- Davidson N.C. and Evans P.R. 1986. The role and potential of man-made and man-modified wetlands in the enhancement of the survival of overwintering shorebirds. *Colonial Waterbirds* 9 (2): 176-188.
- Davidson N.C. and Evans P.R. 1987. Habitat reconstruction and creation: its role and potential in the conservation of waders in The Conservation of International Flyway Populations of Waders (Davidson N.C. and Pienkowski M.W., eds.). Wader Study Group 49/IRWB Special Publication 7, pp. 139-145.
- Davidson N.C., Laffoley D. A., Doody J.P., Way L.S., Gordon J., Key R., Pienkowski M.W., Mitchell R. and Duff K. 1991. *Nature Conservation and Estuaries in Great Britain*. Nature Conservation Council, Petersborough.
- Desprez R., Rybarczyk H., Wilson J.G., Ducrottoy P., Sueur F., Olivesi R. and Elkaim B. 1992. Biological impact of eutrophication in the Bay of Somme and the induction and impact of anoxia. *Neth. Sea Res.* 30: 149-159.
- Everett R.A. 1994. Macroalgae in marine soft-sediment communities: effects on benthic faunal assemblages. *J. Exp. Mar. Biol. Ecol.* 175: 253-274.
- Goss-Custard J.D. and Moser M.E. 1988. Rates of change in the numbers of Dunlin *Calidris alpina* wintering in British estuaries in relation to the spread of *Spartina anglica*. *Journ. Appl. Ecol.* 25: 95-109.
- Goss-Custard J.D. and West A.D. 1997. The concept of carrying capacity and shorebirds. In *Effect of habitat loss and change on waterbirds* (Goss-Custard J.D., Rufino R. and Luís A., eds.) ITE Symposium 30/Wetlands International Publication 42, London.
- Goss-Custard J.D., Durell S.E.A. le V. dit, Clarke R.T., Beintema A.J., Caldow R.W.G., Meininger P.L. and Smit C.J. 1996a. Population dynamics: predicting the consequences of habitat change at the continental scale. In *The Oystercatcher. From individuals to populations* (Goss-Custard J.D., ed.), pp. 352-381. Oxford University Press, Oxford.
- Goss-Custard J.D., West, A.D., Clarke R.T., Caldow R.W.G. and Durell S.E.A. le V. dit. 1996b. The carrying capacity of coastal habitats for Oystercatchers. In *The Oystercatcher. From individuals to populations* (Goss-Custard J.D., ed.), pp. 327-351. Oxford University Press, Oxford.
- Goss-Custard J.D., West, A.D., Clarke R.T., Caldow R.W.G. and Durell S.E.A. le V. dit. 1996c. How Oystercatchers survive the winter. In *The Oystercatcher. From individuals to populations* (Goss-Custard J.D., ed.), pp. 133-154. Oxford University Press, Oxford.
- Hötter H. 1994. Wadden Sea birds and embankments - can artificial wetlands compensate for losses due to land claims? *Ophelia Suppl.* 6: 279-295.
- Lambeck R.H., Goss-Custard J.D. and Triplet P. 1996. Oystercatchers and the man in the coastal zone. In *The Oystercatcher. From individuals to populations* (Goss-Custard J.D., ed.), pp. 289-326. Oxford University Press, Oxford.
- Lopes R., Cabral J., Múrias T. and Marques J.C. 1998. Contribuição para o conhecimento da dieta do Pilrito-comum *Calidris alpina* e da Tarambola-cinzenta *Pluvialis squatarola* no estuário do Mondego. *Airo* 9 (1/2): 28-33.
- Marques J.C., Rodrigues L.S.B. and Nogueira A.J.A. 1993a. Intertidal macrobenthic communities structure in the Mondego estuary (Western Portugal): reference situation. *Vie Milieu* 43: 177-187.
- Marques J.C., Maranhão P. and Pardal M.A. 1993b. Human impact assessment on the subtidal macrobenthic community structure in the Mondego estuary (Western Portugal). *Estuar. Coast. Shelf Sci.* 37: 403-419.
- Meire P.M., Schekkman H. and Meininger P. 1994. Consumption of benthic invertebrates by waterbirds in the Oosterschelde estuary, SW Netherlands. *Hydrobiologia* 282/283: 525-546.
- Múrias T. 1997. Effects of habitat loss on waders (Aves, Charadrii) in the Mondego estuary (Portugal). PhD Thesis, University of Coimbra. Coimbra.

- Múrias T., Cabral J.A., Marques J.C. and Goss-Custard J.D. 1996. Short-term effects of intertidal macroalgae blooms on the microhabitat selection and feeding behaviour of wading birds in the Mondego estuary (West Portugal). *Estuar. Coast. Shelf Sci.* 43: 677-688.
- Múrias T., Cabral J.A., Lopes R. and Carlos Marques J.C. 2002. Effects of eutrophication on waders (Aves: Charadrii) in the Mondego estuary: a multi-level approach (Chapter of this book).
- Neves R. and Rufino R. 1995. Importância ornitológica das salinas: o caso particular do estuário do Sado. *Estudos de Biologia e Conservação da Natureza* 15. ICN, Lisboa.
- Perez-Hurtado A. and Hortas F. 1991. Information about the use of the salines and fish-ponds by wintering waders in Cadiz Bay, southwest Spain. *Wader Study Group Bull.* 66: 48-53.
- Perez-Hurtado A. and Hortas F. 1993a. Actividad trófica de limícolas invernantes en salinas Y cultivos piscícolas de la Bahía de Cádiz. Doñana, *Acta Vert.* 20 (2): 103-123.
- Perez-Hurtado A. and Hortas F. 1993b. Importancia de la Bahía de Cádiz para las poblaciones de limícolas invernantes e influencia de las transformaciones humanas. *Ardeola* 40 (2): 133-142.
- Raffaelli D.G., Hull S.C. and Milne H. 1989. Long-term changes in nutrients, weed mats and shorebirds in an estuarine system. *Cah. Biol. Mar.* 30: 258-270.
- Raffaelli J.D., Limia J., Hull S. and Pont S. 1991. Interaction between the amphipod *Corophium volutator* and macroalgal mats on estuarine mudflats. *J. Mar. Biol. Assoc. UK.* 71: 899-908.
- Rufino R. and Neves R. 1992. The effect on wader populations of the conversion of salinas into fishfarms. *IWRB special Publication* 20: 177-182.
- Rufino R., Araújo A., Pina J.P. and Miranda P.S. 1984. The use of salinas by waders in Algarve. South Portugal Wader Study Group Bull. 42: 41-42.
- Rehfishch M. 1994. Man-made lagoons and how their attractiveness to waders might be increased by manipulating biomass of an insect benthos. *J. Appl. Ecol.* 31: 383-401.
- Soulsby P.G., Lowthion D. and Houston M. 1982. Effects of macroalgal mats on the ecology of intertidal mudflats. *Mar. Pollut. Bull.* 13: 162-166.
- Smit J.C. and Piersma T. 1989. Numbers, midwinter distribution, and migration of wader populations using the East Atlantic flyway. In *Flyways and Reserve Networks for Waterbirds* (Boyd H. and Pirov J.H., eds.), IWRB Special Publication 9, pp. 24-63.
- Velasquez C.P. and Hockey P.A.R. 1990. Seasonal abundance, habitat selection and energy consumption of waterbirds at the Berg River estuary, South Africa. *Ostrich* 62: 109-123.
- Velasquez C.P. and Hockey P.A.R. 1991. The importance of supratidal foraging habitats for waders in a south temperate estuary. *Ardea* 80: 243-253.

Série

Investigação

•

Coimbra  
Imprensa da Universidade

2002